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# **ATTACHMENT-7-1-3-2-AIR EMISSIONS IMPACT ASSESSMENT FOR EPA** LICENCE APPLICATION

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Date of Issue

24 March 2022

**Document History** 

Document Reference		Original Issue Date	
CN/217501.1057AR0	1	24 March 2022	
Revision Level	Revision Date	Description	Sections Affected

**Record of Approval** 

Details	Written by	Approved by
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Date	24 March 2022	24 March 2022
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#### **EXECUTIVE SUMMARY**

This report presents the assessment of air quality impacts as a result of the operation of the Amazon Data Services Ireland Ltd. ("ADSIL" or 'the applicant') data storage facility (the subject 'installation' under this licence application) located at Greenhills Road, Tallaght, Dublin 24. The site is occupied by three no. data storage facility buildings, termed Buildings A, B and C, along with ancillary elements.

The installation requires a continuous supply of electricity to operate. During normal operations, the facility is supplied electricity from the national grid. Outside of normal operations, the facility is first supplied electricity by some or all of the onsite battery installations and then by some or all of the onsite backup generators. Outside of routine testing and maintenance, the operation of these back-up generators is typically only required under the following emergency circumstances:

- A loss, reduction or instability of grid power supply,
- Critical maintenance to power systems.
- A request from the utility supplier (or third party acting on its behalf) to reduce grid electricity load

The air dispersion modelling has been carried out using the United States Environmental Protection Agency's regulated model AERMOD. The AERMOD model has USEPA regulatory status and is one of the advanced models recommended within the air modelling guidance document 'Air Dispersion Modelling from Industrial Installations Guidance Note (AG4)' published by the EPA in Ireland. The modelling of air emissions is carried out to assess concentrations of nitrogen dioxide (NO2) at a variety of locations beyond the site boundary. The modelling assessment also included the current impact of additional facilities with  $NO_X$  emissions in the vicinity of the site.

The assessment has been undertaken for the installation which includes Buildings A, Building A Extension, Building B and Building C, the installation has 37 no. 5.19 MW<sub>th</sub> diesel powered back-up generators, 9 no. 6.6 MW<sub>th</sub> diesel powered back-up generators, 4 no. fire pump generators (0.42 MW<sub>th</sub>). The 4 no diesel powered fire pump generators have been scoped out of this air modelling assessment as it is not expected that they would cause any significant impacts on ambient air quality considering their smaller scale (compared to the data hall back-up generators) and the low number required for use at any one time.

In total, the air dispersion modelling includes 46 no. emergency back-up generators on the installation campus. Building A and its Extension have 24 no. 5.19 MW $_{th}$  back-up generators, Building B has 13 no. 5.19 MW $_{th}$  back-up generators and Building C has 9 no. 6.6 MW $_{th}$  back-up generators.

The assessment has determined the ambient air quality impact of the site and any air quality constraints that may be present. The diesel generators will be used solely for emergency operation (i.e. less than 500 hours per year) and thus the emission limit values outlined in the Medium Combustion Plant Directive are not applicable to the diesel generators on site.

A number of modelling scenarios are investigated for the purposes of this assessment. Both normal day-to-day testing operations are considered as well as emergency operations. Normal testing operations involve the diesel generators operating for up to 30 minutes on a weekly basis at 25% load with no more than one generator tested at the same time. Quarterly maintenance testing of the generators at 90% for 1-hour individually was also included in the modelling assessment. Emergency operation is based on 72 emergency hours modelled according to the USEPA methodology.

#### **Cumulative Air Emissions**

A cumulative impact assessment of the installation and nearby sites within a 1km radius was also conducted. Sites which hold an IED licence from the EPA were assessed for relevant air emissions. There are 2 no. IED licenced sites within 1 km of the subject site, Bimeda Animal Health Limited (Licence No. P0357-01) and Microprint (Licence No. P0659-02). However, these sites do not have licenced  $NO_X$  emission points and therefore there is no potential for cumulative impact and they have not been considered within the cumulative assessment.

Additionally, the Applicant operates a separate data storage facility approximately 940m to the west of the subject site, on the junction of Airton and Belgard Road, Tallaght. Because the Applicant has sufficient information about the emissions associated with emergency back-up generator testing, maintenance and emergency operations at this second facility, these emission sources have been included in the cumulative assessment.

## Assessment of The Facility

The NO<sub>2</sub> modelling results at the worst-case location at and beyond the site boundary are based on the operation of 42 of the 46 no. back-up diesel generations for 72 hours per year using the USEPA methodology outlined within the guidance document titled 'Additional Clarification Regarding Application of Appendix W Modelling Guidance for the 1-Hour National Ambient Air Quality Standard' as well as considering scheduled weekly testing and quarterly maintenance testing of all 46 no. back-up generators from the installation.

The results indicate that the ambient ground level concentrations are in compliance with the relevant air quality standards for NO<sub>2</sub>. For the worst-case year, emissions from the site lead to an ambient NO<sub>2</sub> concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8 percentile) and 81% of the annual limit value at the worst-case off-site receptor.

The UK EA assessment methodology determined that in any year, the diesel generators can run for 672 hours before there is a likelihood of an exceedance at the nearest residential receptor (at a 98th percentile confidence level).

#### **Cumulative Assessment**

The cumulative assessment involved modelling the emergency operation of 42 of the 46 no. standby generators and scheduled testing of the 46 no. standby generators on the subject site in addition to the emissions from the emergency operation and scheduled testing of the standby generators on a separate site to the west owned by the Applicant.

The results indicate that the ambient ground level concentrations are in compliance with the relevant air quality standards for  $NO_2$ . For the worst-case year, emissions from the site lead to an ambient  $NO_2$  concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8th percentile) and 82% of the annual limit value at the worst-case off-site receptor.

The UK EA assessment methodology determined that in any year, the diesel generators can run for 63 hours before there is a likelihood of an exceedance at the nearest residential receptor (at a 98<sup>th</sup> percentile confidence level).

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#### Conclusion

In summary, emissions to atmosphere of NO<sub>2</sub>, as the main polluting substance (as defined in the Schedule of EPA (Industrial Emissions) (Licensing) Regulations 2013, S.I. No. 137 of 2013) from the standby generators, will be in compliance with the ambient air quality standards which are based on the protection of the environment and human health. Therefore, no significant impacts to the ambient air quality environment are predicted.

In terms of impacts in nearby ecologically sensitive areas, the ambient level of  $NO_X$  in the Dodder Valley pNHA (site code 000991), c. 850m to the south-east, due to emissions from the subject site, will be a small fraction of the ambient air quality standard for the protection of vegetation. The nearest European site is Glenasmole Valley SAC c. 4.1 km south of the site. The increased distance of this designated area from the subject site means that it is highly unlikely that airborne pollution could affect any European site.

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#### 1.0 INTRODUCTION

This report presents the assessment of air quality impacts as a result of the operation of the Amazon Data Services Ireland Ltd. ("ADSIL" or 'the applicant') data storage facility (the subject 'installation' under this licence application) located at Greenhills Road, Tallaght, Dublin 24. The site is occupied by three no. data storage installation buildings, termed Building A, Building B and Building C, along with ancillary elements.

The installation requires a continuous supply of electricity to operate. During normal operations, the facility is supplied electricity from the national grid. Outside of normal operations, the facility is first supplied electricity by some or all of the onsite battery installations and then by some or all of the onsite backup generators. Outside of routine testing and maintenance, the operation of these back-up generators is typically only required under the following emergency circumstances:

- A loss, reduction or instability of grid power supply,
- Critical maintenance to power systems,
- A request from the utility supplier (or third party acting on its behalf) to reduce grid electricity load.

The air dispersion modelling has been carried out using the United States Environmental Protection Agency's regulated model AERMOD<sup>(1)</sup>. The AERMOD model has USEPA regulatory status and is one of the advanced models recommended within the air modelling guidance document 'Air Dispersion Modelling from Industrial Installations Guidance Note (AG4)' published by the EPA in Ireland<sup>(2)</sup>. The modelling of air emissions is carried out to assess concentrations of nitrogen dioxide (NO<sub>2</sub>) at a variety of locations beyond the site boundary. The assessment has been undertaken for Buildings A, Building A Extension, Building B and Building C. Building A has 20 no. back-up generators with the Extension having 4 no. back-up generators. Building B has 13 no. back-up generators and Building C has 9 no. back-up generators. In total, the air dispersion modelling includes 46 no. back-up diesel generators on the campus.

The assessment has determined the ambient air quality impact of the site and any air quality constraints that may be present. The diesel generators will be used solely for emergency operation (i.e. less than 500 hours per year) and thus the emission limit values outlined in the Medium Combustion Plant Directive are not applicable to the diesel generators on site.

A number of modelling scenarios are investigated for the purposes of this assessment. Both normal day-to-day testing operations are considered as well as emergency operations. Normal testing operations involve the diesel generators operating for 30 minutes on a weekly basis at 25% load with no more than one generator tested at the same time. Quarterly maintenance testing of the generators at 90% for 1-hour individually was also included in the modelling assessment. Emergency operation is based on 72 emergency hours modelled according to the USEPA methodology.

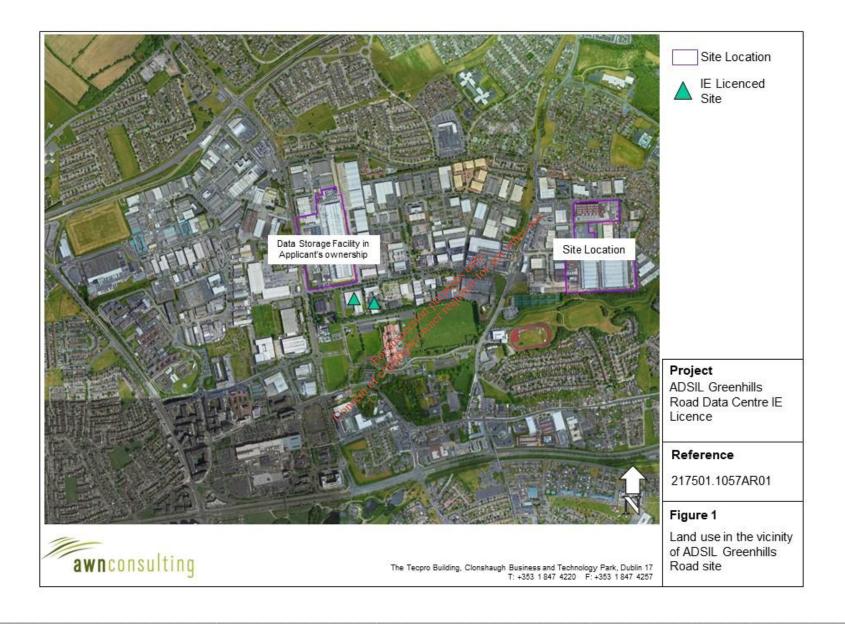
A cumulative impact assessment of the installation and nearby sites within a 1km radius was also conducted. Sites which hold an IED licence from the EPA were assessed for relevant air emissions. There are 2 no. IED licenced sites within 1 km of the subject site, Bimeda Animal Health Limited (Licence No. P0357-01) and Microprint (Licence No. P0659-02). However, these sites do not have licenced NO<sub>X</sub> emission points and therefore there is no potential for cumulative impact and they have not been considered within the cumulative assessment.

Additionally, the Applicant operates a separate data storage facility approximately 940m to the west of the subject site, on the junction of Airton and Belgard Road,

Tallaght. Because the Applicant has sufficient information about the emissions associated with emergency back-up generator testing, maintenance and emergency operations at this second facility, these emission sources have been included in the cumulative assessment. The location of the subject site, the IED licenced site and separate data storage facility to the west are shown below in Figure 1.

Information supporting the conclusions is detailed in the following sections. The assessment methodology and study inputs are presented in Section 2 and Section 3. Background pollutant concentrations are summarised in Section 4. The process emissions and modelling inputs for on-site plant are presented in Section 5. The dispersion modelling results are presented in Section 6 and the assessment summaries are presented in Section 7. The model formulation is detailed in Appendix I and a review of the meteorological data used is detailed in Appendix II.

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## 2.0 ASSESSMENT CRITERIA

## 2.1 Ambient Air Quality Standards

In order to reduce the risk to health from poor air quality, national and European statutory bodies have set limit values in ambient air for a range of air pollutants. These limit values or "Air Quality Standards" are health or environmental-based levels for which additional factors may be considered. The applicable standards in Ireland include the Air Quality Standards Regulations 2011, implement the obligations under Directive 2008/50/EC of the European Parliament and of the Council of 21 May 2008 on ambient air quality and cleaner air for Europe (see Table 1). The ambient air quality standards applicable for NO<sub>2</sub> are outlined in this Directive. Directive 2008/50/EC has also set limit values for other pollutants including SO<sub>2</sub>. Levels of SO<sub>2</sub> emitted from the installation will be insignificant and have been screened out of the assessment.

Air quality significance criteria are assessed on the basis of compliance with the appropriate standards or limit values. The standards outlined in Table 1 have been used in the current assessment to determine the potential impact of NO<sub>2</sub> emissions from the installation on ambient air quality.

Table 1Ambient Air Quality Standards

Pollutant	Regulation Note 1	Limit Type on the standard of	Value
Nitrogen Dioxide	2008/50/EC	Hourly limit for protection of human health - not to be exceeded more than 18 times/year	200 μg/m³
(1002)		Annual limit for protection of human health	40 μg/m³
Nitrogen Oxides (NO <sub>x</sub> )	2008/50/EC	Critical level for protection of vegetation	30 μg/m³ NO + NO <sub>2</sub>
Sulphur Dioxide (SO <sub>2</sub> )	2008/50/EC	Critical level for protection of vegetation (Annual & Winter)	20 μg/m³ SO <sub>2</sub>

Note 1 EU 2008/50/EC – Clean Air For Europe (CAFÉ) Directive replaces the previous Air Framework Directive (1996/30/EC) and daughter directives 1999/30/EC and 2000/69/EC

#### 2.2 Industrial Emissions Directive and Medium Combustion Plant Directive

The Industrial Emissions Directive (IED) (Directive 2010/75/EU) was adopted on 7 January 2013 and is the key European Directive which covers the regulation of the majority of processes in the EU. As part of the IED Article 15, paragraph 2, requires that Emissions Limit Values (ELVs) are based on best available techniques (BAT) and the relevant sector Reference Document of Best Available Techniques (BREF documents).

The most relevant BAT sector document for the activities at the installation is the Best Available Techniques (BAT) Reference Document for Large Combustion Plants LCP. There are no ELVs set out in the LCP BAT that are applicable to the individual emergency back-up generators.

The individual emergency back-up generators are 5.19 MW<sub>th</sub> for Buildings A and B, and 6.6 MW<sub>th</sub> for Building C, therefore the Medium Combustion Plant (MCP)

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Regulations (S.I No. 595 of 2017), which transposed the Medium Combustion Plant Directive ((EU) 2015/2193) is a relevant consideration in respect of the individual plant.

The installation requires a continuous supply of electricity to operate. During normal operations, the facility is supplied electricity from the national grid. Outside of normal operations, the facility is first supplied electricity by some or all of the onsite battery installations and then by some or all of the onsite backup generators. Outside of routine testing and maintenance, the operation of these back-up generators is typically only required under the following emergency circumstances:

- A loss, reduction or instability of grid power supply,
- Critical maintenance to power systems,
- A request from the utility supplier (or third party acting on its behalf) to reduce grid electricity load.

The diesel generators are for emergency back-up only and are not anticipated to operate in excess of 500 hours per annum. Therefore, the emergency generators as proposed are exempt from complying with the relevant ELVs set out in the MCP Directive subject to Section 13(3) of the Medium Combustion Plant (MCP) Regulations.

The UK Environment Agency assessment methodology in Section 6.2 below determined that the standby generators could operate for 672 hours before there is a likelihood of an exceedance of the ambient air quality standard (at a 98<sup>th</sup> percentile confidence level). However, the UK guidance recommends that there should be no running time restrictions placed on standby generators which provide power on site only during an emergency power outage.

## 2.3 Sensitive areas or areas of special interest

An Appropriate Assessment (AA) Screening Report (Attachment 6-3-4) has been prepared by Moore Group and has been submitted as part of the licence application for the site.

The AA Screening identifies that the lands in which the installation is located have no formal designations, and that the nearest European site to the Installation is the Glenasmole Valley SAC (Site Code 001209) located over 4.14 km to the south.

Emissions of  $NO_X$  have the potential to impact vegetation and sensitive plant species. Directive 2008/50/EC has set limit values for vegetation effects as per Table 1. As such it is typical to assess the impact of  $NO_X$  emissions from a facility on any nearby sensitive ecological areas in close proximity to the site. There are no European sites within 1 km of the subject site as noted above.

The closest sensitive ecological area is the Dodder Valley Proposed Natural Heritage Area (pNHA) (site code 000991) which is located c. 850 m south-east of the subject site. Dispersion modelling of  $NO_X$  emissions from the installation has been conducted within the Dodder Valley pNHA to determine the potential impact to vegetation as a result of emissions from the back-up generators on site. Emissions from the back-up generators are not predicted to be significant at this distance from the installation as emission concentrations peak at the site boundary and fall off rapidly with increasing distance from the installation.

#### 3.0 ASSESSMENT METHODOLOGY

Emissions from the installation are modelled using the AERMOD dispersion model (Version 21112) which has been developed by the U.S. Environmental Protection Agency (USEPA)<sup>(1)</sup> and following guidance issued by the EPA<sup>(2)</sup>. The model is a steady-state Gaussian plume model used to assess pollutant concentrations associated with industrial sources and has replaced ISCST3<sup>(4)</sup> as the regulatory model by the USEPA for modelling emissions from industrial sources in both flat and rolling terrain<sup>(5-7)</sup>. The model has more advanced algorithms and gives better agreement with monitoring data in extensive validation studies<sup>(8-11)</sup>. An overview of the AERMOD dispersion model is outlined in Appendix I.

The air dispersion modelling input data consisted of information on the physical environment (including building dimensions and terrain features), design details from all emission points on-site and five years of appropriate hourly meteorological data. Using this input data the model predicted ambient ground level concentrations beyond the site boundary for each hour of the modelled meteorological years. The model post-processed the data to identify the location and maximum of the worst-case ground level concentration. This worst-case concentration is then added to the background concentration to give the worst-case predicted environmental concentration (PEC). The PEC is then compared with the relevant ambient air quality standard to assess the significance of the releases from the site.

The modelling aims to achieve compliance with the guidance outlined within the EPA AG4 Guidance for Air Dispersion Modelling hroughout this study a worst-case approach was taken. This will most likely lead to an over-estimation of the levels that will arise in practice. The worst-case assumptions are outlined below:

- Maximum predicted concentrations are reported in this study, even if no residential receptors are reported in this maximum;
- Conservative background concentrations are used in the assessment;
- The effects of building downwash, due to on-site buildings, are included in the model;
- Emergency operations were assumed to occur for a maximum of 72 hours per year calculated according to USEPA methodology, in reality generators are likely to be used for maintenance and testing purposes only.

## 3.1 Air Dispersion Modelling Methodology

The United States Environmental Protection Agency (USEPA) approved AERMOD dispersion model has been used to predict the ground level concentrations (GLC) of compounds emitted from the principal emission sources on-site.

The modelling incorporated the following features:

• Two receptor grids are included at which concentrations are modelled. Receptors are mapped with sufficient resolution to ensure all localised "hotspots" are identified without adding unduly to processing time. The receptor grids are based on Cartesian grids with the site at the centre. The outer grid measures 5 x 5 km with the site at the centre and with concentrations calculated at 100m intervals. The inner grid measures 1 x 1 km with the site at the centre and with concentrations calculated at 50m intervals. Boundary receptor locations are also placed along the boundary of the site, at 25m intervals, giving a total of 3,142 calculation points for the model.

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 Discrete receptors are also added to the model to represent nearby residential receptors. Receptors were modelled at a height of 1.8m to represent breathing height.

- All on-site buildings and significant process structures are mapped into the
  computer to create a three dimensional visualisation of the site and its emission
  points. Buildings and process structures can influence the passage of airflow
  over the emission stacks and draw plumes down towards the ground (termed
  building downwash). The stacks themselves can influence airflow in the same
  way as buildings by causing low pressure regions behind them (termed stack
  tip downwash). Both building and stack tip downwash are incorporated into the
  modelling.
- Detailed terrain has been mapped into the model using SRTM data with 30m resolution. The site is located in an area of complex terrain. All terrain features have been mapped in detail into the model using the terrain pre-processor AERMAP<sup>(12)</sup>.
- Hourly-sequenced meteorological information has been used in the model.
   Meteorological data over a five year period (Casement Aerodrome 2017 2021) is used in the model (see Figure 2 and Appendix II).
- The source and emissions data, including stack dimensions, gas volumes and emission temperatures have been incorporated into the model.

## 3.2 Terrain

The AERMOD air dispersion model has a terrain pre-processor AERMAP<sup>(12)</sup> which is used to map the physical environment in detail over the receptor grid. The digital terrain input data used in the AERMAP pre-processor is obtained from SRTM. This data is run to obtain for each receptor point the terrain height and the terrain height scale. The terrain height scale is used in AERMOD to calculate the critical dividing streamline height, H<sub>crit</sub>, for each receptor. The terrain height scale is derived from the Digital Elevation Model (DEM) files in AERMAP by computing the relief height of the DEM point relative to the height of the receptor and determining the slope. If the slope is less than 10%, the program goes to the next DEM point. If the slope is 10% or greater, the controlling hill height is updated if it is higher than the stored hill height.

In areas of complex terrain, AERMOD models the impact of terrain using the concept of the dividing streamline ( $H_c$ ). As outlined in the AERMOD model formulation<sup>(1)</sup> a plume embedded in the flow below  $H_c$  tends to remain horizontal; it might go around the hill or impact on it. A plume above  $H_c$  will ride over the hill. Associated with this is a tendency for the plume to be depressed toward the terrain surface, for the flow to speed up, and for vertical turbulent intensities to increase.

AERMOD model formulation states that the model "captures the effect of flow above and below the dividing streamline by weighting the plume concentration associated with two possible extreme states of the boundary layer (horizontal plume and terrainfollowing). The relative weighting of the two states depends on: 1) the degree of atmospheric stability; 2) the wind speed; and 3) the plume height relative to terrain. In stable conditions, the horizontal plume "dominates" and is given greater weight while in neutral and unstable conditions, the plume traveling over the terrain is more heavily weighted"<sup>(1)</sup>.

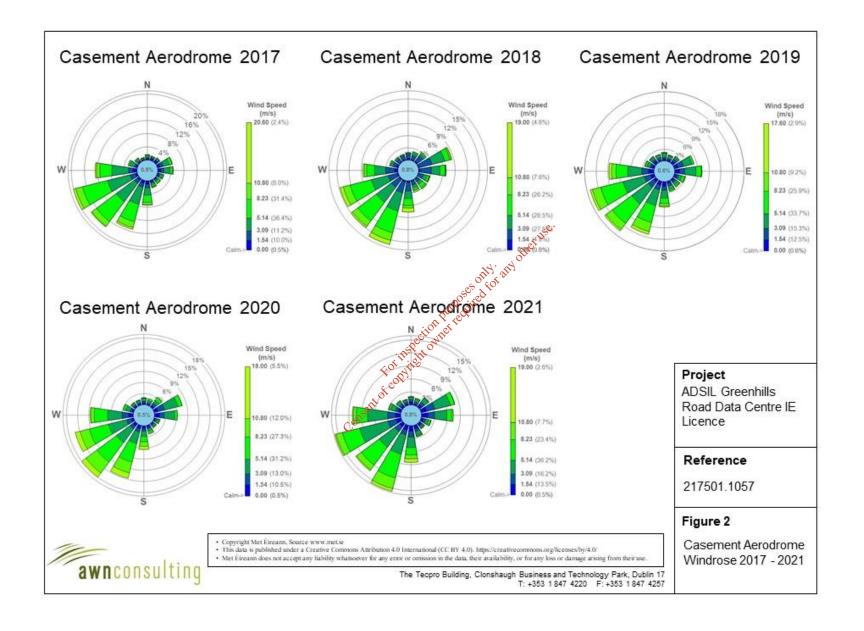
## 3.3 Meteorological Data

The selection of the appropriate meteorological data has followed the guidance issued by the USEPA<sup>(1)</sup>. A primary requirement is that the data used should have a data capture of greater than 90% for all parameters. Casement Aerodrome meteorological station, which is located approximately 6 km west of the site, collects data in the correct

format and has a data collection of greater than 90%. Long-term hourly observations at Casement Aerodrome meteorological station provide an indication of the prevailing wind conditions for the region (see Figure 2 and Appendix II) $^{(13)}$ . Results indicate that the prevailing wind direction is predominantly south-westerly in direction over the period 2017 – 2021. The mean wind speed is approximately 5.5 m/s over the period 2017 - 2021.

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## 3.4 Geophysical Considerations

AERMOD simulates the dispersion process using planetary boundary layer (PBL) scaling theory<sup>(1)</sup>. PBL depth and the dispersion of pollutants within this layer are influenced by specific surface characteristics such as surface roughness, albedo and the availability of surface moisture. Surface roughness is a measure of the aerodynamic roughness of the surface and is related to the height of the roughness element. Albedo is a measure of the reflectivity of the surface whilst the Bowen ratio is a measure of the availability of surface moisture.

AERMOD incorporates a meteorological pre-processor AERMET<sup>(14)</sup> to enable the calculation of the appropriate parameters. The AERMET meteorological preprocessor requires the input of surface characteristics, including surface roughness ( $z_0$ ), Bowen Ratio and albedo by sector and season, as well as hourly observations of wind speed, wind direction, cloud cover, and temperature. The values of albedo, Bowen Ratio and surface roughness depend on land-use type (e.g., urban, cultivated land etc) and vary with seasons and wind direction. The assessment of appropriate land-use type is carried out to a distance of 10km from the meteorological station for Bowen Ratio and albedo and to a distance of 1km for surface roughness in line with USEPA recommendations<sup>(14,15)</sup> as outlined in Appendix II.

In relation to AERMOD, detailed guidance for calculating the relevant surface parameters has been published<sup>(15)</sup>. The most pertinent features are:

- The surface characteristics should be those of the meteorological site (Casement Aerodrome) rather than the installation;
- Surface roughness should use a default 1km radius upwind of the meteorological tower and should be based on an inverse-distance weighted geometric mean. If land use varies around the site, the land use should be subdivided by sectors with a minimum sector size of 30°;
- Bowen ratio and albedoshould be based on a 10km grid. The Bowen ratio should be based on an un-weighted geometric mean. The albedo should be based on a simple un-weighted arithmetic mean.

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AERMOD has an associated pre-processor, AERSURFACE<sup>(15)</sup> which has representative values for these parameters depending on land use type. The AERSURFACE pre-processor currently only accepts NLCD92 land use data which covers the USA. Thus, manual input of surface parameters is necessary when modelling in Ireland. Ordnance survey discovery maps (1:50,000) and digital maps such as those provided by the EPA, National Parks and Wildlife Service (NPWS) and Google Earth® are useful in determining the relevant land use in the region of the meteorological station. The Alaska Department of Environmental Conservation has issued a guidance note for the manual calculation of geometric mean for surface roughness and Bowen ratio for use in AERMET<sup>(16)</sup>. This approach has been applied to the current site with full details provided in Appendix II.

## 3.5 Building Downwash

When modelling emissions from an industrial installation, stacks which are relatively short can be subjected to additional turbulence due to the presence of nearby buildings. Buildings are considered nearby if they are within five times the lesser of the building height or maximum projected building width (but not greater than 800m).

The USEPA has defined the "Good Engineering Practice" (GEP) stack height as the building height plus 1.5 times the lesser of the building height or maximum projected

building width. It is generally considered unlikely that building downwash will occur when stacks are at or greater than GEP<sup>(17)</sup>.

When stacks are less than this height, building downwash will tend to occur. As the wind approaches a building it is forced upwards and around the building leading to the formation of turbulent eddies. In the lee of the building these eddies will lead to downward mixing (reduced plume centreline and reduced plume rise) and the creation of a cavity zone (near wake) where re-circulation of the air can occur. Plumes released from short stacks may be entrained in this airflow leading to higher ground level concentrations than in the absence of the building.

The Plume Rise Model Enhancements (PRIME)<sup>(9,10)</sup> plume rise and building downwash algorithms, which calculates the impact of buildings on plume rise and dispersion, have been incorporated into AERMOD. The building input processor BPIP-PRIME produces the parameters which are required in order to run PRIME. The model takes into account the position of each stack relative to each relevant building and the projected shape of each building for 36 wind directions (at 10° intervals). The model determines the change in plume centreline location with downwind distance based on the slope of the mean streamlines and coupled to a numerical plume rise model<sup>(10)</sup>.

Given that the stacks are less than 2.5 times the lesser of the building height or maximum projected building width, building downwash was taken into account and the PRIME algorithm run prior to modelling with AERMOD. The dominant building for each relevant stack will vary as a function of wind direction and relative building heights.

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#### 4.0 BACKGROUND CONCENTRATIONS OF POLLUTANTS

Air quality monitoring programs have been undertaken in recent years by the EPA and Local Authorities<sup>(18,19)</sup>. The most recent annual report on air quality "Air Quality in Ireland 2020"<sup>(19)</sup>, details the range and scope of monitoring undertaken throughout Ireland. As part of the implementation of the Framework Directive on Air Quality (1996/62/EC), four air quality zones have been defined in Ireland for air quality management and assessment purposes<sup>(18)</sup>. Dublin is defined as Zone A and Cork as Zone B. Zone C is composed of 23 towns with a population of greater than 15,000. The remainder of the country, which represents rural Ireland but also includes all towns with a population of less than 15,000 is defined as Zone D. In terms of air monitoring, Greenhills Road is categorized as Zone A<sup>(18)</sup>.

In 2020 the EPA reported<sup>(19)</sup> that Ireland was compliant with EU legal air quality limits at all locations, however this was largely due to the reduction in traffic due to Covid-19 restrictions. The EPA report details the effect that the Covid-19 restrictions had on air monitoring stations, which included reductions of up to 50% at some monitoring stations which have traffic as a dominant source. The report also notes that CSO figures show that while traffic volumes are still slightly below 2019 levels, they have significantly increased since 2020 levels. 2020 concentrations are therefore predicted to be an exceptional year and not consistent with long-term trends. For this reason, they have not been included in the baseline section and previous data has been used to determine the baseline air quality in the region of the site.

With regard to NO<sub>2</sub>, continuous monitoring data from the EPA<sup>(18,19)</sup>, at suburban Zone A background locations in Rathmines, Swords and Ballyfermot show that current levels of NO<sub>2</sub> are below both the annual and 1 four limit values, with annual average levels ranging from 15 - 22  $\mu$ g/m³ in 2019 (see Table 2). Sufficient data is available for all stations to observe long-term trends since 2015<sup>(18)</sup>, with annual average results ranging from 13 – 22  $\mu$ g/m³. The 1-hour concentrations, measured as a 99.8<sup>th</sup> percentile are also in compliance with the 1-hour limit value of 200  $\mu$ g/m³. Based on these results, an estimate of the current background NO<sub>2</sub> concentration in the region of the facility is 17  $\mu$ g/m³.

In relation to the annual average background, the ambient background concentration is added directly to the process concentration with the short-term peaks assumed to have an ambient background concentration of twice the annual mean background concentration.

The methodology for converting  $NO_X$  to  $NO_2$  is based on the ozone limiting method (OLM) approach based on an initial  $NO_2/NO_X$  ratio of 0.1 and a background ozone level of 55  $\mu$ g/m³ based on a review of EPA data for similar Zone A locations<sup>(19)</sup>.

**Table 2** Annual Mean and 99.8<sup>th</sup> Percentile 1-Hour NO₂ Concentrations In Zone A Locations (µg/m³)

Station	Averaging Deried	Year					
Station	Averaging Period	2015	2016	2017	2018	2019	
Rathmines	Annual Mean NO <sub>2</sub> (μg/m <sup>3</sup> )	18	20	17	20	22	
Rathmines	99.8 <sup>th</sup> %ile 1-hr NO <sub>2</sub> (μg/m <sup>3</sup> )	105	88	86	87	102	
Swords	Annual Mean NO <sub>2</sub> (µg/m <sup>3</sup> )	13	16	14	16	15	
Swords	99.8 <sup>th</sup> %ile 1-hr NO <sub>2</sub> (μg/m <sup>3</sup> )	93	96	79	85	80	
Dallyformat	Annual Mean NO <sub>2</sub> (µg/m <sup>3</sup> )	16	17	17	17	20	
Ballyfermot	99.8 <sup>th</sup> %ile 1-hr NO <sub>2</sub> (μg/m <sup>3</sup> )	127	90	112	101	101	



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#### 5.0 PROCESS EMISSIONS

The installation has no major emissions to air and only has potential (emergency generators) emissions that will generate quantities of air pollutants listed as a Principal Pollution Substance (S.I. No. 137/2013 - Environmental Protection Agency (Industrial Emissions) (Licensing) Regulations 2013).

The data storage facility has 46 no. back-up diesel generators in total. Building A has 20 no. back-up generators and Building A Extension has 4 no. back-up generators. Building B has 13 no. back-up generators and Building C has 9 no. back-up generators. Each generator has an associated stack the heights of which were designed in an iterative fashion to provide for adequate dispersion of pollutants. The stacks associated with Building A are primarily 6 m horizontal stacks for 14 no. generators, with 6 no. generators having 8 m vertical stacks. Building A Extension has vertical 18 m stacks on all 4 no. generators. The stacks associated with Building B are 16.7 m and vertical. Building C has 21.7 m vertical stacks. All stacks are measured relative to ground level.

Two of the back-up diesel generators associated with Building A and one back-up generator associated with Building B and C have been modelled as "catcher" generators to provide redundancy to the remaining generators on site. Therefore, in the event of a power failure at the site 42 of the 46 no. backup generators will be operational.

Modelling for  $NO_2$  was undertaken in detail. In relation to CO,  $PM_{10}$  and  $PM_{2.5}$  no detailed modelling was undertaken. Emissions of these pollutants are significantly lower than the  $NO_X$  emissions from the generators relative to their ambient air quality standards and thus ensuring compliance with the  $NO_2$  ambient limit value will ensure compliance for all other pollutants. For example, the emission of CO from the generators is at least eight times lower than  $NO_X$  whilst the CO ambient air quality standard is  $10,000~\mu g/m^3$  compared to the 1-hour  $NO_2$  standard of  $200~\mu g/m^3$ . Similarly, levels of  $PM_{10}/PM_{2.5}$  emitted from the generators will be significantly lower than  $NO_X$  emissions whilst the ambient air quality standards are comparable.

In addition to the 5.19 MW<sub>th</sub> and 6.6 MW<sub>th</sub> emergency back-up generators which will power the site in the event of a power failure, the site also includes 4 no. fire pump generators (0.42 MW<sub>th</sub>) generators. The diesel powered fire pump generators have been scoped out of this air modelling assessment as it is not expected that they would cause any significant impacts on ambient air quality considering their smaller scale (compared to the stationary data hall back-up generators) and the low number required for use at any one time.

Two testing regimes for the back-up generators have also been included in the model as outlined below, all testing was assumed to occur from 8am to 5pm, Monday to Friday only.

- Test 1: Testing once per week of all 46 no. back-up generators on the campus at 25% load for a maximum of 30 minutes each, one generator at a time, sequentially.
- Test 2: All 46 no. back-up generators will be periodically tested individually at 90% load for a maximum of four hours per year. This is incorporated into the dispersion model as one generator operating, at 90% load, for one full hour, once per week during every quarter (assumed to be January, April, June and October for the purpose of this assessment).

The modelling has considered maintenance testing of the back-up generators once per week during four months of the year, however, in reality maintenance testing will occur for a maximum of 4 hours per year or only once every quarter for each back-up generators. Thus, the worst-case approach used in this study will lead to an overestimation of the actual levels that will arise.

## 5.1 Emergency Operations

The diesel generators will operate in an emergency scenario when there is a power outage in the area. In addition, testing of the generators will be required as outlined above.

There are two methodologies used to determine the impact from the operation of the diesel generators on ambient air quality. Both methodologies from the USEPA and UK Environment Agency have been used in this assessment, this follows the guidance outlined in Appendix K of the Irish EPA document AG4<sup>(2)</sup>. Emission details can be seen in Table 4.

USEPA Guidance suggests that for emergency operations, an average hourly emission rate should be used rather than the maximum hourly rate<sup>(3)</sup>. As a result, the maximum hourly emission rates from the diesel generators are reduced by  $\frac{72}{8760}$  and the diesel generators are modelled over a period of one-full year.

A second methodology has been published by the UK Environment Agency. The consultation document is entitled "Diese Generator Short-Term NO2 Impact Assessment"(20). The methodology is based on considering the statistical likelihood of an exceedance of the NO<sub>2</sub> hourly limit value (18 exceedances are allowable per year before the air standard is deemed to have been exceeded). The assessment assumes a hypergeometric distribution to assess the likelihood of exceedance hours coinciding with the emergency operational bours of the diesel generators. The cumulative hypergeometric distribution of 19 and more hours per year is computed and the probability of an exceedance determined. The guidance suggests that the 95th percentile confidence levelshould be used to indicate if an exceedance is likely. More recent guidance(21) has recommended this probability should be multiplied by a factor of 2.5 and therefore the 98th percentile confidence level should be used to indicate if an exceedance is likely. The guidance suggests that the assessment should be conducted at the nearest residential receptor or at locations where people are likely to be exposed and that there should be no running time restrictions on these generators when providing power on site during an emergency.

Both the methodology advised in the USEPA guidance as well as the approach described in the UK EA guidance have been applied for the emergency scenario modelled in this study to ensure a robust assessment of predicted air quality impacts from the diesel generators. This also follows the guidance outlined in Appendix K of the EPA AG4 guidance<sup>(2)</sup>.

The modelling is undertaken to assess the impact to ambient air quality from the following two emergency operations scenarios:

**The Facility:** This includes the emergency operation of 42 no. of the 46 no. diesel generators (the remaining 4 no. generators serving as a "catcher" generator for Buildings A - C). The scenario also included testing of all 46 no. generators as described above. The process emissions are outlined in Table 4;

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Cumulative Assessment: A cumulative impact assessment of the facility and nearby sites within a 1km radius was also conducted. Sites which hold an IED licence from the EPA were assessed for relevant air emissions. There are 2 no. IED licenced site within 1 km of the subject site, Bimeda Animal Health Limited (Licence No. P0357-01) and Microprint (Licence No. P0659-02). However, these sites do not have licenced  $NO_{\rm X}$  emission points and therefore there is no potential for cumulative impact and they have not been considered within the cumulative assessment. Additionally, the Applicant operates a separate data storage facility approximately 940m to the west of the subject site, on the junction of Airton and Belgard Road, Tallaght. Because the Applicant has sufficient information about the emissions associated with emergency back-up generator testing, maintenance and emergency operations at this second facility, these emission sources have been included in the cumulative assessment.

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 Table 3
 Summary of Process Emission Information for all Buildings Associated with the Facility

			Cross- Sectional Area (m²)		Volume Flow (Nm <sup>3</sup> /hr at 15% Ref. O <sub>2</sub> )	Exit Velocity (m/sec actual)	NO <sub>x</sub>	
Stack Reference	Stack Height Above Ground Level (m)	Exit Diameter (m)		Temp (K)			Concentration (mg/Nm³ at 15% Ref. O <sub>2</sub> )	Mass Emission (g/s)
Emergency Operation and Testing of Back-up Diesel Generators, Building A & B (90% load)	6m & 8m - Building A 18m - Building A Extension 16.7m - Building B	0.325	0.08	744.65	15,946	88.64	673	0.024 Note 1 / 2.98 Note 2
Testing of Diesel Generators, Building A & B (25% load)				619.05	4,516	32.73	847	0.531 Note 3
Emergency Operation and Testing of Back-up Diesel Generators, Building C (90% load)	21.7	0.6	0.00	741.15 othe	18,160	27.68	857	0.036 Note 1 / 4.32 Note 2
Testing of Diesel Generators, Building C (25% load)		21.7		5716.05	6,777	12.96	417	0.393 Note 3

Note 1 Reduced emission rates based on USEPA protocol (assuming 72 hours / annum) used to model emissions during emergency operation of generators (90% load)

Maximum emission rates for diesel generators (based on 90% load) used to model emissions during emergency operation of generators for UK EA assessment methodology and for quarterly testing for USEPA assessment methodology

Note 3 Emission rates used to model emissions during scheduled testing at 25% load conducted once per week.

#### 6.0 RESULTS

## 6.1 Emergency Operations (USEPA Methodology)

The NO<sub>2</sub> modelling results at the worst-case location at and beyond the site boundary are detailed in Table 4 based on the operation of 42 of the 46 no. back-up diesel generations for 72 hours per year using the USEPA methodology outlined within the guidance document titled 'Additional Clarification Regarding Application of Appendix W Modelling Guidance for the 1-Hour National Ambient Air Quality Standard' (3) as well as considering scheduled weekly testing and quarterly maintenance testing of all 46 no. back-up generators from the installation.

The results indicate that the ambient ground level concentrations are within the relevant air quality standards for NO<sub>2</sub>. For the worst-case year modelled, emissions from the site lead to an ambient NO<sub>2</sub> concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8<sup>th</sup> percentile) and 81% of the annual limit value at the worst-case off-site receptor. Concentrations decrease with distance from the site boundary. The geographical variations in the 1-hour mean (99.8<sup>th</sup> percentile) and annual mean NO<sub>2</sub> ground level concentrations for the Facility Scenario are illustrated as concentration contours in Figures 3 and 4. The locations of the maximum concentrations for NO<sub>2</sub> are close to the boundary of the site with concentrations decreasing with distance from the installation.

**Table 4** Dispersion Model Results for Nitrogen Dioxide (NO<sub>2</sub>) – Emergency Operations & Scheduled Testing

Pollutant / Year	Background (µg/m³)	Averaging Period	Process Contribution (µg/m³)	Predicted Environmental Concentration (µg/m³)	Standard (µg/m³) Note 1
NO <sub>2</sub> / 2017	34	99.8th%ile of 1-Hr Means	121.4	155.4	200
	17 (	Annual mean	15.4	32.4	40
NO <sub>2</sub> / 2018	34	99.8th%ile of 1-Hr Means	124.7	158.7	200
	17	Annual mean	14.2	31.2	40
NO <sub>2</sub> / 2019	34	99.8th%ile of 1-Hr Means	111.0	145.0	200
	17	Annual mean	13.5	30.5	40
NO <sub>2</sub> / 2020	34	99.8th%ile of 1-Hr Means	101.1	135.1	200
	17	Annual mean	13.8	30.8	40
NO <sub>2</sub> / 2021	34	99.8th%ile of 1-Hr Means	91.9	125.9	200
	17	Annual mean	13.2	30.2	40

Note 1 Air Quality Standards 2011 (from EU Directive 2008/50/EC and S.I. 180 of 2011)

## 6.2 Emergency Operations (UK EA Methodology)

Emissions from 42 of the 46 no. standby generators were assessed using the UK Environment Agency methodology. The methodology, based on considering the statistical likelihood of an exceedance of the NO<sub>2</sub> hourly limit value assuming a hypergeometric distribution, has been undertaken at the worst-case residential receptor for the Facility Scenario. The methodology, based on considering the

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statistical likelihood of an exceedance of the  $NO_2$  hourly limit value assuming a hypergeometric distribution, has been undertaken at the worst-case residential / sensitive receptor. The cumulative hypergeometric distribution of 19 and more hours per year is computed and the probability of an exceedance determined. The results have been compared to the  $98^{th}$  percentile confidence level to indicate if an exceedance is likely at various operational hours for the diesel generators. The results (Table 5 and Figure 5) indicate that in the worst-case year, the diesel generators can operate for the 672 hours per year before there is a likelihood of an exceedance of the ambient air quality standard (at a  $98^{th}$  percentile confidence level).

**Table 5** Hypergeometric Statistical Results at Worst-case Residential Receptor – Emergency Operations Scenario

Pollutant / Year / Scenario	Hours of operation (Hours) (98 <sup>th</sup> %ile) Allowed Prior To Exceedance Of Limit Value	UK Guidance – Probability Value = 0.02 (98 <sup>th</sup> %ile) <sup>Note 1</sup>
NO <sub>2</sub> / 2017	955	
NO <sub>2</sub> / 2018	830	
NO <sub>2</sub> / 2019	672	0.02
NO <sub>2</sub> / 2020	830	
NO <sub>2</sub> / 2021	762	

Note 1 Guidance Outlined In UK Environment Agency (2019) Emissions from specified generators - Guidance on dispersion modelling for oxides of nitrogen assessment from specified generators

## 6.3 Cumulative Assessment (USEPA Methodology)

The cumulative assessment involved modelling the emergency operation of 42 of the 46 no. standby generators and scheduled testing of the 46 no. standby generators on the subject site in addition to the emissions from a second data storage facility owned by the Applicant to the west of the site. The cumulative NO<sub>2</sub> modelling results at the worst-case location at and beyond the site boundary are detailed in Table 6.

The results indicate that the ambient ground level concentrations are within the relevant air quality standards for NO<sub>2</sub>. For the worst-case year modelled, emissions from the site lead to an ambient NO<sub>2</sub> concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8<sup>th</sup> percentile) and 82% of the annual limit value at the worst-case off-site receptor. Concentrations decrease with distance from the site boundary. The geographical variations in the 1-hour mean (99.8<sup>th</sup> percentile) and annual mean NO<sub>2</sub> ground level concentrations for the Cumulative Scenario are illustrated as concentration contours in Figures 6 and 7. The locations of the maximum concentrations for NO<sub>2</sub> are close to the boundary of the site with concentrations decreasing with distance from the installation.

**Table 6** Dispersion Model Results for Nitrogen Dioxide (NO<sub>2</sub>) – Cumulative Assessment

Pollutant / Year	Background (µg/m³)	Averaging Period	Process Contribution (µg/m³)	Predicted Environmental Concentration (µg/m³)	Standard (µg/m³) Note 1
NO <sub>2</sub> / 2017	34	99.8th%ile of 1-Hr Means	121.4	155.4	200
	17	Annual mean	15.9	32.9	40
NO <sub>2</sub> / 2018		99.8th%ile of 1-Hr Means	124.7	158.7	200
	17	Annual mean	14.5	31.5	40
NO <sub>2</sub> / 2019	34	99.8th%ile of 1-Hr Means	111.0	145.0	200
	17	Annual mean	13.9	30.9	40
NO <sub>2</sub> / 2020	34	99.8th%ile of 1-Hr Means	101.1	135.1	200
	17	Annual mean	14.3	31.3	40
NO <sub>2</sub> / 2021	34	99.8th%ile of 1-Hr Means	91.9	125.9	200
Note 1	17	Annual mean	13.7 et 155	30.7	40

Note 1 Air Quality Standards 2011 (from EU Directive 2008/50/EC and S.I. 180 of 2011)

## 6.4 Cumulative Assessment (UK EA Methodology)

The cumulative assessment using the UK Environment Agency methodology involved the emergency operation of 42 of the 46 no. standby generators on the subject site and emissions from the standby generators on a separate site to the west owned by the Applicant. The methodology based on considering the statistical likelihood of an exceedance of the NO<sub>2</sub> hourly limit value assuming a hypergeometric distribution, has been undertaken at the worst-case residential receptor for the Cumulative Scenario. The methodology, based on considering the statistical likelihood of an exceedance of the NO<sub>2</sub> hourly limit value assuming a hypergeometric distribution, has been undertaken at the worst-case residential / sensitive receptor. The cumulative hypergeometric distribution of 19 and more hours per year is computed and the probability of an exceedance determined. The results have been compared to the 98th percentile confidence level to indicate if an exceedance is likely at various operational hours for the diesel generators. The results indicate that in the worst-case year, the diesel generators can operate for the 63 hours per year before there is a likelihood of an exceedance of the ambient air quality standard (at a 98th percentile confidence level) (see Table 7).

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Table 7	Hypergeometric Statistical Results at Worst-case Residential Receptor
<ul> <li>Cumulative</li> </ul>	Assessment

Pollutant / Year / Scenario	Hours of operation (Hours) (98 <sup>th</sup> %ile) Allowed Prior To Exceedance Of Limit Value	UK Guidance – Probability Value = 0.02 (98 <sup>th</sup> %ile) <sup>Note 1</sup>	
NO <sub>2</sub> / 2017	110		
NO <sub>2</sub> / 2018	63		
NO <sub>2</sub> / 2019	84	0.02	
NO <sub>2</sub> / 2020	67		
NO <sub>2</sub> / 2021	68		

Outlined In UK Environment Agency (2019) Emissions from specified generators Guidance on dispersion modelling for oxides of nitrogen assessment from specified generators

## 6.5 Air Quality Impact on European Sites

The modelling of air emissions from the site was carried out to assess the concentrations of nitrogen dioxide  $(NO_x)$  beyond the site boundary at the nearest sensitive ecological area. There are no designated European sites within 1 km of the installation (see Section 2.3). The modelling undertaken has been reviewed and compared with the limit values in Table 1.

The nearest ecologically sensitive area to the subject site is the Dodder Valley Proposed Natural Heritage Area (pNHA) (site code 000991) which is located c. 850 m south-east of the subject site.

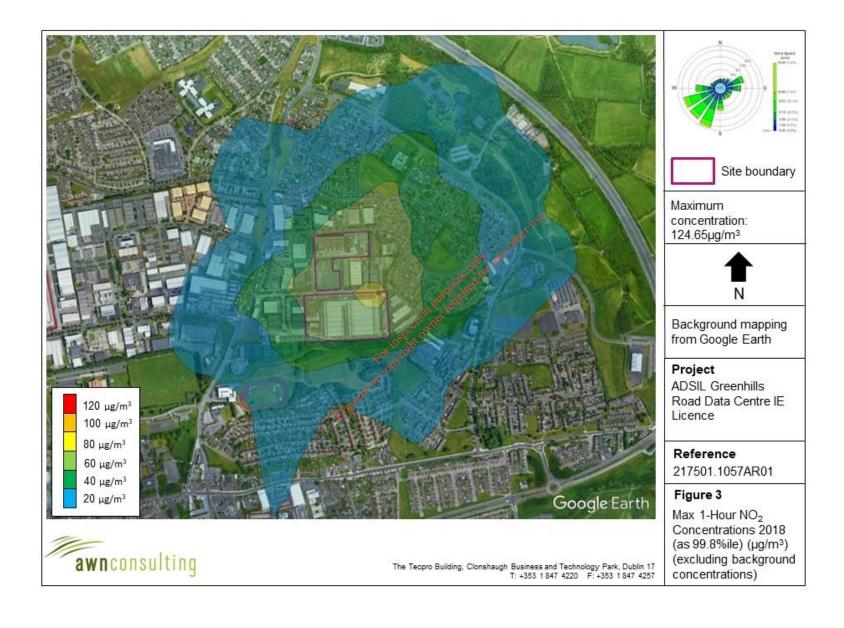
The results shown in Table 8 indicate that the ambient ground level concentrations at the nearest ecologically sensitive area are within the relevant air quality standard for  $NO_X$ . For the worst-case year the modelled process emissions from the site lead to an ambient  $NO_X$  concentration which is at most 0.5% of the annual limit value at the worst-case ecological receptor as shown in Table 8.

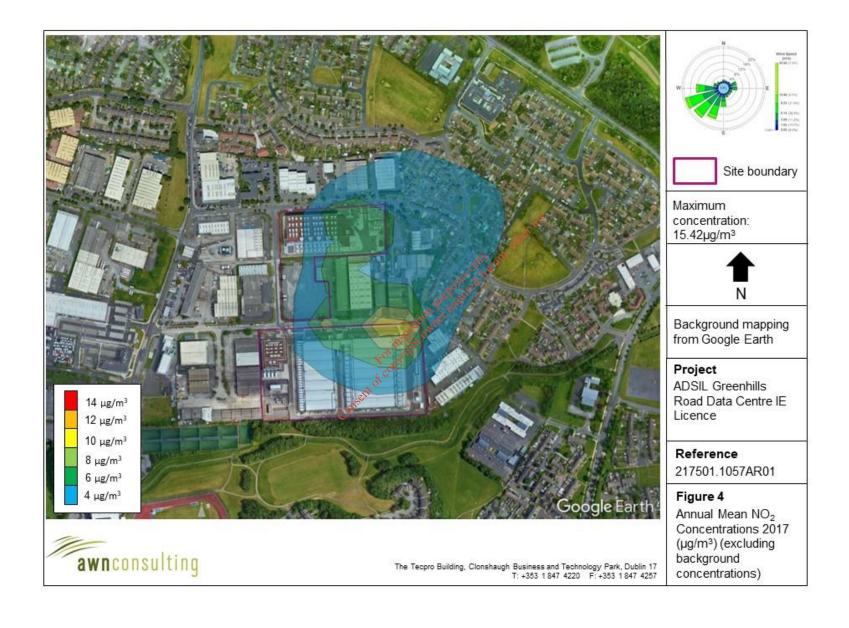
**Table 8** NO<sub>x</sub> Dispersion Model Results In Ecologically Sensitive Areas

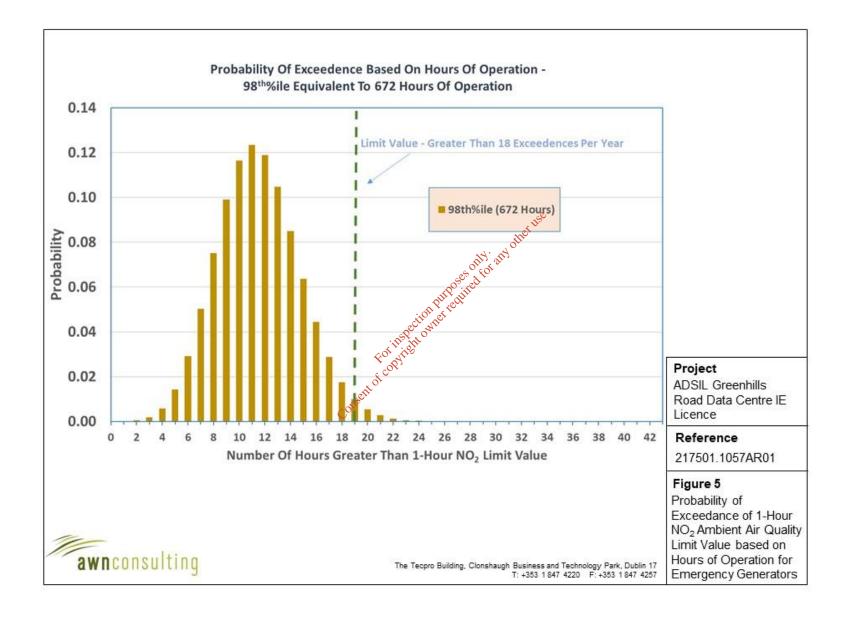
Pollutant / Year	Averaging Period	Process Contribution NO <sub>x</sub> (µg/m³)	NO <sub>x</sub> Limit Value (μg/m³)	NO <sub>x</sub> As % Of Limit Value
NO <sub>x</sub> / 2017	Annual Mean	0.12	30	0.39%
NO <sub>x</sub> / 2018	Annual Mean	0.10	30	0.34%
NO <sub>x</sub> / 2019	Annual Mean	0.14	30	0.47%
NO <sub>x</sub> / 2020	Annual Mean	0.11	30	0.38%
NO <sub>x</sub> / 2021	Annual Mean	0.15	30	0.50%

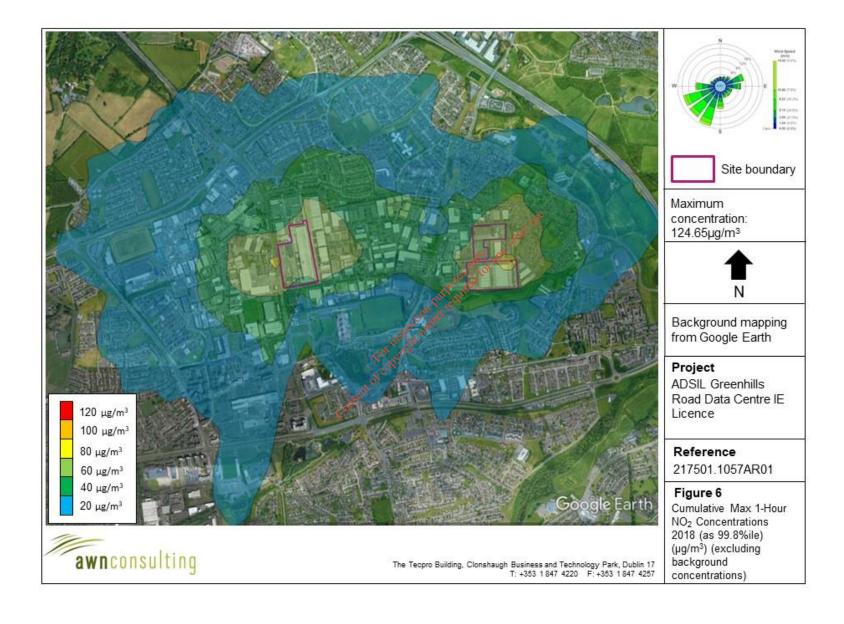
In terms of impacts in the nearby ecologically sensitive areas, the ambient level of  $NO_X$  in the Dodder Valley pNHA (site code 000991); c. 850m south-east, due to emissions during operations, will be a small fraction of the ambient air quality standard for the protection of vegetation.

The nearest designated European site to the installation is the Glenasmole Valley SAC; c. 4.1km south. Based on the separation distance, and dispersion of pollutant from the site it is highly unlikely that airborne pollution from the installation would affect the nearest or any other European site.











#### 7.0 ASSESSMENT SUMMARY

The assessment was carried out the determine the ambient air quality impact of the site. As the diesel generators will be used solely for emergency operation (i.e. less than 500 hours per year) thus the emission limit values outlined in the Medium Combustion Plant Directive are not applicable to the diesel generators on site.

## The Facility

The NO<sub>2</sub> modelling results at the worst-case location at and beyond the site boundary are based on the operation of 42 of the 46 no. back-up diesel generations for 72 hours per year using the USEPA methodology outlined within the guidance document titled 'Additional Clarification Regarding Application of Appendix W Modelling Guidance for the 1-Hour National Ambient Air Quality Standard' as well as considering scheduled weekly testing and quarterly maintenance testing of all 46 no. back-up generators from the installation.

The results indicate that the ambient ground level concentrations are in compliance with the relevant air quality standards for NO<sub>2</sub>. For the worst-case year, emissions from the site lead to an ambient NO<sub>2</sub> concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8th percentile) and 81% of the annual limit value at the worst-case off-site eceptor.

The UK EA assessment methodology determined that in any year, the diesel generators can run for 672 hours before there is a likelihood of an exceedance at the nearest residential receptor (at a 98<sup>th</sup> percentile confidence level).

#### **Cumulative Assessment**

The cumulative assessment involved modelling the emergency operation of 42 of the 46 no. standby generators and scheduled testing of the 46 no. standby generators on the subject site in addition to the emissions from the emergency operation and scheduled testing of the standby generators on a separate site to the west owned by the Applicant.

The results indicate that the ambient ground level concentrations are in compliance with the relevant air quality standards for NO<sub>2</sub>. For the worst-case year, emissions from the site lead to an ambient NO<sub>2</sub> concentration (including background) which is 79% of the maximum ambient 1-hour limit value (measured as a 99.8th percentile) and 82% of the annual limit value at the worst-case off-site receptor.

The UK EA assessment methodology determined that in any year, the diesel generators can run for 63 hours before there is a likelihood of an exceedance at the nearest residential receptor (at a 98<sup>th</sup> percentile confidence level).

## Conclusion

In summary, emissions to atmosphere of NO<sub>2</sub>, as the main polluting substance (as defined in the Schedule of EPA (Industrial Emissions) (Licensing) Regulations 2013, S.I. No. 137 of 2013) from the standby generators, will be in compliance with the ambient air quality standards which are based on the protection of the environment and human health. Therefore, no significant impacts to the ambient air quality environment are predicted.

In terms of impacts in the nearby ecologically sensitive areas, the ambient level of  $NO_X$  in the Dodder Valley pNHA (site code 000991), c. 850m to the south-east, due to

emissions from the subject site, will be a small fraction of the ambient air quality standard for the protection of vegetation. The nearest European site is Glenasmole Valley SAC c. 4.1 km south of the site. The increased distance of this designated area from the subject site means that it is highly unlikely that airborne pollution could affect any European site.

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## 8.0 REFERENCES

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#### APPENDIX I

## **Description of the AERMOD Model**

The AERMOD dispersion model has been developed in part by the U.S. Environmental Protection Agency (USEPA)<sup>(1,5)</sup>. The model is a steady-state Gaussian model used to assess pollutant concentrations associated with industrial sources. The model is an enhancement on the Industrial Source Complex-Short Term 3 (ISCST3) model which has been widely used for emissions from industrial sources.

Improvements over the ISCST3 model include the treatment of the vertical distribution of concentration within the plume. ISCST3 assumes a Gaussian distribution in both the horizontal and vertical direction under all weather conditions. AERMOD with PRIME, however, treats the vertical distribution as non-Gaussian under convective (unstable) conditions while maintaining a Gaussian distribution in both the horizontal and vertical direction during stable conditions. This treatment reflects the fact that the plume is skewed upwards under convective conditions due to the greater intensity of turbulence above the plume than below. The result is a more accurate portrayal of actual conditions using the AERMOD model. AERMOD also enhances the turbulence of night-time urban boundary layers thus simulating the influence of the urban heat island.

In contrast to ISCST3, AERMOD is widely applicable in all types of terrain. Differentiation of the simple versus complex terrain is unnecessary with AERMOD. In complex terrain, AERMOD employs the dividing-streamline concept in a simplified simulation of the effects of plume-terrain interactions. In the dividing-streamline concept, flow below this height remains horizontal, and flow above this height tends to rise up and over terrain. Extensive validation studies have found that AERMOD (precursor to AERMOD with PRIME) performs better than ISCST3 for many applications and as well or better than CTDMPLUS for several complex terrain data sets<sup>(7)</sup>.

Due to the proximity to surrounding buildings, the PRIME (Plume Rise Model Enhancements) building downwash algorithm has been incorporated into the model to determine the influence (wake effects) of these buildings on dispersion in each direction considered. The PRIME algorithm takes into account the position of the stack relative to the building in calculating building downwash. In the absence of the building, the plume from the stack will rise due to momentum and/or buoyancy forces. Wind streamlines act on the plume leads to the bending over of the plume as it disperses. However, due to the presence of the building, wind streamlines are disrupted leading to a lowering of the plume centreline.

When there are multiple buildings, the building tier leading to the largest cavity height is used to determine building downwash. The cavity height calculation is an empirical formula based on building height, the length scale (which is a factor of building height & width) and the cavity length (which is based on building width, length and height). As the direction of the wind will lead to the identification of differing dominant tiers, calculations are carried out in intervals of 10 degrees.

In PRIME, the nature of the wind streamline disruption as it passes over the dominant building tier is a function of the exact dimensions of the building and the angle at which the wind approaches the building. Once the streamline encounters the zone of influence of the building, two forces act on the plume. Firstly, the disruption caused by the building leads to increased turbulence and enhances horizontal and vertical dispersion. Secondly, the streamline descends in the lee of the building due to the reduced pressure and drags the plume (or part of) nearer to the ground, leading to higher ground level concentrations. The model calculates the descent of the plume as a function of the building shape and, using a numerical plume rise model, calculates the change in the plume centreline location with distance downwind.

The immediate zone in the lee of the building is termed the cavity or near wake and is characterised by high intensity turbulence and an area of uniform low pressure. Plume mass captured by the cavity region is re-emitted to the far wake as a ground-level volume source. The volume source is located at the base of the lee wall of the building, but is only evaluated near the end of the near wake and beyond. In this region, the disruption caused by the building downwash gradually fades with distance to ambient values downwind of the building.

AERMOD has made substantial improvements in the area of plume growth rates in comparison to ISCST3<sup>(4,8)</sup>. ISCST3 approximates turbulence using six Pasquill-Gifford-Turner Stability Classes and bases the resulting dispersion curves upon surface release experiments. This treatment, however, cannot explicitly account for turbulence in the formulation. AERMOD is based on the more realistic modern planetary boundary layer (PBL) theory which allows turbulence to vary with height. This use of turbulence-based plume growth with height leads to a substantial advancement over the ISCST3 treatment.

Improvements have also been made in relation to mixing height<sup>(4,8)</sup>. The treatment of mixing height by ISCST3 is based on a single morning upper air sounding each day. AERMOD, however, calculates mixing height on an hourly basis based on the morning upper air sounding and the surface energy balance, accounting for the solar radiation, cloud cover, reflectivity of the ground and the latent heat due to evaporation from the ground cover. This more advanced formulation provides a more realistic sequence of the diurnal mixing height changes.

AERMOD also has the capability of modelling both unstable (convective) conditions and stable (inversion) conditions. The stability of the atmosphere is defined by the sign of the sensible heat flux. Where the sensible heat flux is positive the atmosphere is unstable whereas when the sensible heat flux is negative the atmosphere is defined as stable. The sensible heat flux is dependent on the net radiation and the available surface moisture (Bowen Ratio). Under stable (inversion) conditions, AERMOD has specific algorithms to account for plume rise under stable conditions, mechanical mixing heights under stable conditions and vertical and lateral dispersion in the stable boundary laver.

AERMOD also contains improved algorithms for dealing with low wind speed (near calm) conditions. As a result, AERMOD can produce model estimates for conditions when the wind speed may be less than 1 m/s, but still greater than the instrument threshold.

# APPENDIX II Meteorological Data - AERMET

AERMOD incorporates a meteorological pre-processor AERMET<sup>(14)</sup>. AERMET allows AERMOD to account for changes in the plume behaviour with height. AERMET calculates hourly boundary layer parameters for use by AERMOD, including friction velocity, Monin-Obukhov length, convective velocity scale, convective (CBL) and stable boundary layer (SBL) height and surface heat flux. AERMOD uses this information to calculate concentrations in a manner that accounts for changes in dispersion rate with height, allows for a non-Gaussian plume in convective conditions, and accounts for a dispersion rate that is a continuous function of meteorology.

The AERMET meteorological preprocessor requires the input of surface characteristics, including surface roughness  $(z_0)$ , Bowen Ratio and albedo by sector and season, as well as hourly observations of wind speed, wind direction, cloud cover, and temperature. A morning sounding from a representative upper air station, latitude, longitude, time zone, and wind speed threshold are also required.

Two files are produced by AERMET for input to the AERMOD dispersion model. The surface file contains observed and calculated surface variables, one record per hour. The profile file contains the observations made at each level of a meteorological tower, if available, or the one-level observations taken from other representative data, one record level per hour.

From the surface characteristics (i.e. surface roughness albedo and amount of moisture available (Bowen Ratio)) AERMET calculates several boundary layer parameters that are important in the evolution of the boundary layer, which, in turn, influences the dispersion of pollutants. These parameters include the surface friction velocity, which is a measure of the vertical transport of horizontal momentum; the sensible heat flux, which is the vertical transport of heat to/from the surface; the Monin-Oburhov length which is a stability parameter relating the surface friction velocity to the sensible heat flux; the daytime mixed layer height; the nocturnal surface layer height and the convective velocity scale which combines the daytime mixed layer height and the sensible heat flux. These parameters all depend on the underlying surface.

The values of albedo, Bowen Ratio and surface roughness depend on land-use type (e.g., urban, cultivated land etc) and vary with seasons and wind direction. The assessment of appropriate land-use types is carried out in line with USEPA recommendations<sup>(5)</sup> and using the detailed methodology outlined by the Alaska Department of Environmental Conservation<sup>(16)</sup>. AERMET has also been updated to allow for an adjustment of the surface friction velocity (u\*) for low wind speed stable conditions based on the work of Qian and Venkatram (BLM, 2011). Previously, the model had a tendency to over-predict concentrations produced by near-ground sources in stable conditions.

## Surface Roughness

Surface roughness length is the height above the ground at which the wind speed goes to zero. Surface roughness length is defined by the individual elements on the landscape such as trees and buildings. In order to determine surface roughness length, the USEPA recommends that a representative length be defined for each sector, based on geometric mean of the inverse distance area-weighted land use within the sector, by using the eight land use categories outlined by the USEPA. The area-weighted surface roughness length derived from the land use classification within a radius of 1km from Casement Aerodrome is shown in Table A1.

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**Table A1** Surface Roughness based on an inverse distance area-weighted average of the land use within a 1km radius of Casement Aerodrome

Sector	Area Weighted Land Use Classification	Spring	Summer	Autumn	Winter <sup>Note 1</sup>
0-360	100% Grassland	0.050	0.100	0.010	0.010

Note 1:

Winter defined as periods when surfaces covered permanently by snow whereas autumn is defined as periods when freezing conditions are common, deciduous trees are leafless and no snow is present (lqbal (1983)). Thus for the current location autumn more accurately defines "winter" conditions at the proposed facility.

#### Albedo

Noon-time Albedo is the fraction of the incoming solar radiation that is reflected from the ground when the sun is directly overhead. Albedo is used in calculating the hourly net heat balance at the surface for calculating hourly values of Monin-Obuklov length. The area-weighted arithmetic mean albedo derived from the land use classification over a 10km x 10km area centred on Casement Aerodrome is shown in Table A2.

**Table A2** Albedo based on an area-weighted arithmetic mean of the land use over a 10km x 10km area centred on Casement Aerodrome

Area Weighted Land Use Classification	Spring	Summer.	Autumn	Winter <sup>Note1</sup>
0.5% Water, 30% Urban, 0.5% Coniferous Forest 38% Grassland, 19% Cultivated Land	0.155	0.180	0.187	0.187

Note 1: For the current location autumn more accurately defines winter conditions at the proposed facility.

## **Bowen Ratio**

The Bowen ratio is a measure of the amount of moisture at the surface of the earth. The presence of moisture affects the heat balance resulting from evaporative cooling which, in turn, affects the Monin-Obukhov length which is used in the formulation of the boundary layer. The area-weighted geometric mean Bowen ratio derived from the land use classification over a 10km x 10km area centered on Casement Aerodrome is shown in Table A3.

**Table A3** Bowen Ratio based on an area-weighted geometric mean of the land use over a 10km x 10km area centred on Casement Aerodrome

Area Weighted Land Use Classification	Spring	Summer	Autumn	Winter <sup>Note1</sup>
0.5% Water, 30% Urban, 0.5% Coniferous Forest 38% Grassland, 19% Cultivated Land	0.549	1.06	1.202	1.202

Note 1: For the current location autumn more accurately defines "winter" conditions at the proposed facility.